



This is a self-archived – parallel published version of an original article. This version may differ from the original in pagination and typographic details. When using please cite the original.

Taylor & Francis:

This is an Accepted Manuscript version of the following article, accepted for publication in:

LANDSCAPE RESEARCH

E. Seda Arslan, Paulina Nordström, Asko Ijäs, Reija Hietala & Nora Fagerholm (2021): Perceptions of Cultural Ecosystem Services: spatial differences in urban and rural areas of Kokemäenjoki, Finland, *Landscape Research*, DOI: 10.1080/01426397.2021.1907322

10.1080/01426397.2021.1907322

It is deposited under the terms of the Creative Commons Attribution-NonCommercial-NoDerivatives License (<http://creativecommons.org/licenses/by-nc-nd/4.0/>) which permits non-commercial re-use, distribution, and reproduction in any medium, provided the original work is properly cited, and is not altered, transformed, or built upon in any way.

1 **Spatial distribution of perceived cultural ecosystem services in different**
2 **land cover infrastructure types in urban and rural areas – case**
3 **Kokemäki area, Finland**

4 E. Seda Arslan^{1,2*}, Paulina Nordström³, Asko Ijäs⁴, Reija Hietala² and Nora
5 Fagerholm²

6 *¹Süleyman Demirel University, Faculty of Architecture, Department of Landscape*
7 *Architecture, Turkey; sedaarslan@sdu.edu.tr*

8 *²University of Turku, Department of Geology and Geography, Finland;*
9 *seda.arslan@utu.fi; reija.hietala@utu.fi; nora.fagerholm@utu.fi*

10 *³University of Agder, Department of Global Development and Planning;*
11 *paulinaen@uia.no*

12 *⁴University of Turku, Brahea Centre, Finland; asko.ijas@utu.fi; asko.ijas@gmail.com*

13 *Correspondence: sedaarslan@sdu.edu.tr ; sedaarslan@utu.fi

14 **Orcid ID Numbers:**

15 E. Seda Arslan: [0000-0003-1592-5180](https://orcid.org/0000-0003-1592-5180)

16 Paulina Nordström: [0000-0002-4677-3506](https://orcid.org/0000-0002-4677-3506)

17 Asko Ijäs: [0000-0002-3289-9458](https://orcid.org/0000-0002-3289-9458)

18 Reija Hietala: [0000-0001-8096-056X](https://orcid.org/0000-0001-8096-056X)

19 Nora Fagerholm: [0000-0001-5020-0746](https://orcid.org/0000-0001-5020-0746)

20

21

22

23

24

25 **Spatial distribution of perceived cultural ecosystem services in different**
26 **land cover infrastructure types in urban and rural areas – case**
27 **Kokemäki area, Finland**

28 This study aims to identify and evaluate the spatial distribution of Cultural
29 Ecosystem Services (CES) benefits perceived by people in both urban and rural
30 areas. A public participation GIS (PPGIS) approach was applied with local people
31 who responded to an online survey and mapped their important places related to
32 CES benefits in the Kokemäenjoki area. We explore the perceived ecosystem
33 services of the community using different infrastructure types (green, grey, yellow
34 and blue) based on the Corine Land Cover (CLC) classes. We identified spatial
35 patterns of mapped important places using kernel density estimation and related
36 CES benefit associations with the infrastructures using chi-square residuals. We
37 found that CES in urban areas are provided more often when there is more than
38 one type of infrastructure (e.g. grey and green; grey and blue), but grey
39 infrastructures are preferred in urban areas, while blue infrastructures produce
40 more CES benefits in rural areas.

41 Keywords: Ecosystem services; land cover; infrastructures; participatory mapping;
42 PPGIS; Finland

43 **Introduction**

44 Ecosystem service, as a concept, has gained increasing attention among scientists and
45 decision makers since the publication of the Millennium Ecosystem Assessment, which
46 illustrates the complex dependencies between natural environments and human well-
47 being (Mea, 2005). However, ‘ecosystem services’ requires a clear and logical definition
48 (Ahtiainen & Öhman, 2014). Several studies (Berghöfer et al., 2011; Costanza et al.,
49 1997; Daily, 1997; Haines-Young & Potschin, 2013; Mea, 2005) have defined it in the
50 literature as, for example, the ‘benefits that people obtain from an ecosystem’; however,
51 the definition needs to specify the functions of the ecosystem and its effects on human
52 well-being. Hence, more recent research defines ecosystem services as associations of

53 ecosystem function and physical structure with a combination of other related inputs for
54 human well-being (Burkhard et al., 2012, 2018). The Intergovernmental Science-Policy
55 Platform on Biodiversity and Ecosystem Services (IPBES) Regional Assessment in
56 Europe and Central Asia also revealed the dynamic relationships between nature's
57 contributions to people, biodiversity and ecosystems, and their relevance for human
58 quality of life (Martín-López et al., 2018).

59 It is now broadly realised that ecosystem services are co-produced not only by
60 ecosystem functions alone but also by their interactions with local social systems
61 (Eigenbrod, 2016; Meacham et al., 2016; Reyers et al., 2013). This brings attention to
62 assessing ecosystem services in relation to various land uses, which connect to different
63 human activities, for example along the urban–rural gradient, in order to integrate
64 ecosystem services into the planning process. Bonilla-Bedoya et al. (2020) have
65 developed new tools and techniques along the urban–rural gradient by using different
66 metrics and indexes related to building social environment and landscape characteristics.
67 Rural and urban areas are both unique in terms of their ecological, social and
68 environmental features. Rural areas are dominated by natural environments like pastures,
69 agricultural areas, forests and natural green areas, and they typically include only small
70 human settlements. Urban areas are characterised by built-up areas like densely populated
71 settlements and artificial surfaces, but they also include nature, such as urban parks and
72 forests (Gebre & Gebremedhin, 2019; Von Braun, 2007). Researchers and policy makers
73 can overlook the contribution of the urban and rural areas in terms of providing benefits
74 to each other.

75 There is growing interest in measuring ecosystem services in the urban–rural
76 context (Chen et al., 2019; Soto-Montes-de-Oca et al., 2020; Zhu et al., 2017). It is
77 essential to evaluate which ecosystem services are typical for different land use types to

78 further develop ecosystem service-based planning in the urban–rural gradient (Koschke
79 et al., 2012). Several studies have made these evaluations, mainly by using biophysical
80 ecological data (Felipe-Lucia & Comín, 2015; Lilburne et al., 2020) such as land use/land
81 cover or social-cultural data (Zhang & Ramírez, 2019) using survey or land-based
82 approaches with statistical analysis. However, most of these studies address land use
83 changes rather than ecosystem services provided by different land use and land cover
84 types. Burkhard et al. (2009) and Koschke et al. (2012) measure the provision of
85 ecosystem services in different land cover classes: Broadly, mixed and coniferous forests
86 have been shown to provide provisioning, regulating and cultural services. Provisioning
87 services are also provided by different kinds of agricultural areas and coastal landscapes,
88 while regulating services are linked to water bodies. Burkhard et al. (2009) and Koschke
89 et al. (2012) have also noticed that forests could provide multiple services. In their
90 research, provisioning services are mostly provided by plantation areas and pastures.
91 Regulating services are quite common in all land use types except urban areas. Cultural
92 services are also provided in different environments like urban areas, water bodies,
93 grasslands and pastures.

94 Different types of land uses like green areas (such as forests, shrub and/or
95 herbaceous vegetation, and urban green areas) can be characterised with the term
96 ‘infrastructure’. In recent years, green infrastructure has become important in land use
97 sustainability planning (Hansen & Pauleit, 2014). Recently, ecosystem service studies
98 have become related to different infrastructures, especially green infrastructure, which
99 provides ecosystem services that improve the well-being of both nature and people
100 (Dipeolu & Ibem, 2020). In addition to green infrastructure, similar terminology can also
101 be applied to other land cover types to describe their large-scale distribution. Arable
102 lands, permanent crops, pastures and heterogeneous agricultural areas can be identified

103 as yellow infrastructure; water bodies and wetlands as blue; and built-up areas as grey
104 infrastructure, respectively. Blue and green infrastructures can be considered as
105 ecological infrastructures in ecosystem service–related studies, which try to focus on the
106 importance of infrastructures for identifying ecosystems goods and services spatially
107 (Kati & Jari, 2016; Li et al., 2017). Although no research has used the term ‘yellow
108 infrastructure’, it can be applied to represent agricultural areas (Egoh et al., 2011; Lin et
109 al., 2015; Qiu & Turner, 2013). In urban areas, grey infrastructure is studied more than
110 green infrastructure, and some research deals with the integration of green-grey
111 infrastructures to improve well-being (Svendsen et al., 2012). Infrastructure and
112 ecosystem services concepts have the potential to improve landscape planning in urban
113 areas and provide tools for the more holistic assessment of complex interrelations and
114 dynamics of social–ecological systems (Hansen & Pauleit, 2014).

115 However, applying the ecosystem service concept at the landscape level is
116 challenging because particularly the assessment of cultural ecosystem services (CES)
117 such as aesthetic values, recreation and tourism requires more specific data on landscape
118 character and landscape identity, as well as on non-material landscape benefits (Bachi et
119 al., 2020; Bieling & Plieninger, 2013; Burkhard et al., 2009; Kaymaz, 2013; Small et al.,
120 2017). As introduced by the European Landscape Convention, ‘landscape’ refers to ‘an
121 area, as perceived by people, whose character is the result of the action and interaction of
122 natural and/or human factors’ (CoE, 2000). The Convention identifies landscape as an
123 important component to understand all aspects of human life in all areas (Jones, 2007).
124 Analysing landscape structure with its natural and cultural elements is important for
125 bridging the gap between humans and nature; it is an effective way of explaining
126 landscape characteristics related to ecosystems and, hence, to deliver ecosystem services
127 (Ko & Son, 2018; Ramyar, 2019). Therefore, to provide sustainable landscape planning

128 and management, analysing the landscape's physical features together with its social
129 values is crucial (De Groot, 2006; Fagerholm et al., 2019). At this point, participatory
130 research could be a key element for representing people's perceptions on the benefits of
131 ecosystem services in a place-based way.

132 Participatory research aims to represent the knowledge of local communities and
133 improve public involvement (Sieber, 2006; Kenter, 2016; Ridding et al., 2018), and it is
134 a way of collecting perceptions of landscape and ecosystem services (Herlihy & Knapp,
135 2003). Additionally, it aims to aid decision makers in understanding how people perceive
136 different landscape values linked to human well-being (Elwell et al., 2018; Battisti et al.,
137 2019). Participatory approaches generally promote the understanding of the stakeholders'
138 and communities' knowledge, perceptions and preferences about their living area or
139 environment (Fagerholm et al., 2016; Villamor et al., 2014). Participatory mapping has
140 become common in science and practice (Brown, 2013; Brown & Fagerholm, 2015;
141 Brown & Kyttä, 2014a; Fagerholm et al., 2016; Garcia-Martin et al., 2017). Often called
142 Participatory Geographical Information System or Public Participatory Geographical
143 Information System (PGIS/PPGIS), participatory mapping approaches enable the
144 collection of people's perceptions, values and interests in a place-based way. There is
145 increasing academic and governmental interest in using participatory mapping
146 technologies for decision-making and planning processes. In recent years, participatory
147 mapping has also been used to identify and map ecosystem services (Brown, 2012; Brown
148 & Fagerholm, 2015). However, studies have focused either on the rural (Fagerholm et al.,
149 2016; García-Nieto et al., 2015) or urban (Ko & Son, 2018; Rall et al., 2017; Zhang &
150 Ramírez, 2019) context, and there is limited understanding in assessing ecosystem
151 services in the urban–rural gradient or in relation to different infrastructure types
152 (Palomo-Campesino et al., 2018).

153 The main purpose of this study is to identify and evaluate the spatial distribution
154 of CES perceived by local people in both the urban and rural areas of the Kokemäenjoki
155 area (Southwest Finland, Fig 1). Local Kokemäenjoki residents responded to an online
156 survey and mapped their important places related to different CES in their daily
157 landscape. We explore the perceived ecosystem services of the community in relation to
158 different infrastructure types. We aim to produce information on the CES provided by the
159 study area and promote their acknowledgement in landscape planning. The more specific
160 research questions are:

- 161 • What are the most common land cover infrastructure types (blue, green, yellow,
162 grey) in rural and urban areas in the Kokemäki area, and what is the spatial
163 distribution of important places in them?
- 164 • What types of CES benefits are perceived in each infrastructure type across rural
165 and urban areas?

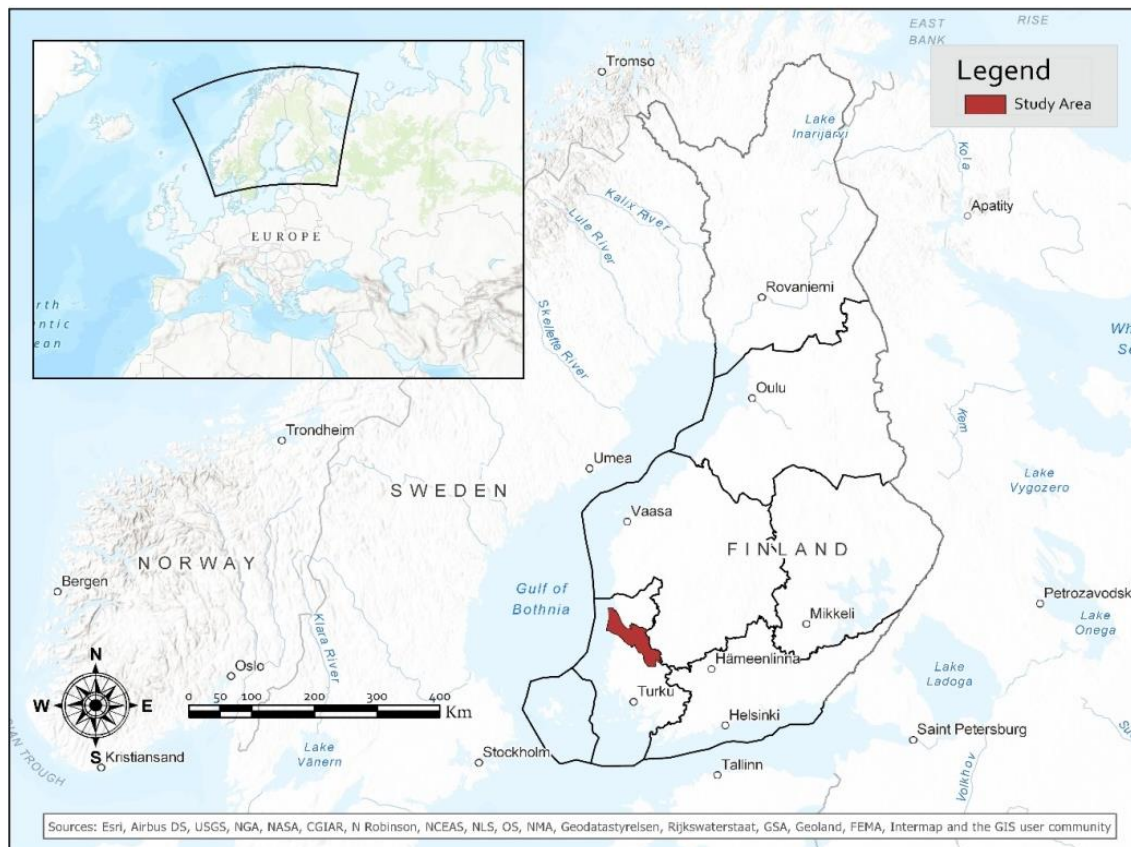
166 Finally, we discuss the relevance of the results for landscape planning and
167 management, particularly in terms of the different infrastructures in rural and urban areas.

168 **Materials and Methods**

169 *2.1 Study area*

170 This study is performed along an urban–rural gradient in the Kokemäki River Basin
171 located in Southwest Finland. Our study area (approximately 2,900 km²) covers a wide
172 array of natural and agricultural environments located along the river catchment. Human
173 population in the river basin is concentrated along the main river channel. The biggest
174 population centres are Pori (84,318 inhabitants), Ulvila (13,009 inhabitants), Harjavalta
175 (7,006 inhabitants), Kokemäki (7,226 inhabitants), and Huittinen (10,075 inhabitants)

176 (OSF, 2020) (Figure 1). Kokemäenjoki, which discharges to the Bothnian Sea, is the
177 longest river for this area. It has been widely used for transportation, irrigation and energy
178 production, and many historical places are also linked to the river. Some parts of the study
179 area have high-level human pressure because of urbanisation and other related reasons
180 such as seasonal tourism destinations along the shoreline.



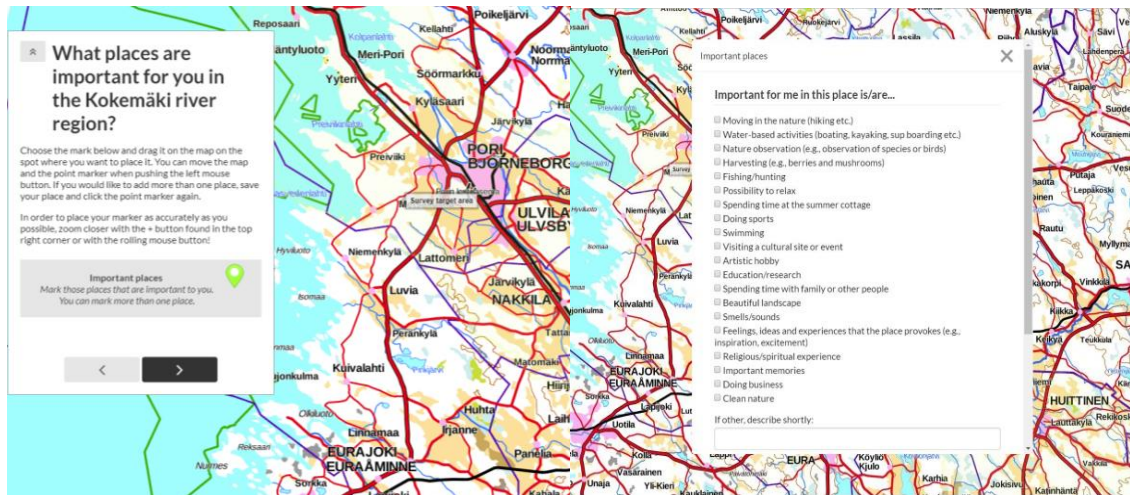
181

182 Figure 1. The Kokemäenjoki area in Southwest Finland.

183 *2.2 Survey contents and data collection*

184 This research was connected to the European Union–funded Interreg Central Baltic
185 Programme 2014–2020 project SustainBaltic (<https://sites.utu.fi/sustainbaltic/>).
186 Perceived ecosystem service benefits were collected through participatory mapping with
187 an online survey (<https://app.maptionnaire.com/en/4424/>) operated on the Maptionnaire
188 platform. Maptionnaire has been widely used for academic research in recent years

189 (Ernoul et al., 2018; Kahila et al., 2017; Kahila-Tani et al., 2019; Lamoureux & Fast,
190 2019). Developed as an online ‘do-it-yourself’ tool, Maptionnaire is an advanced example
191 of PPGIS methodology and allows the mapping of environmental preferences, daily
192 behaviours and knowledge and beliefs to collect spatial data (Kahila-Tani et al., 2019).
193 The survey was open for the month of August 2018, and the survey link was shared via
194 press release in municipalities, regional council actors, regional planners, nature
195 conservation associations and social media (Facebook groups). Additionally, the study
196 area was visited several times to distribute leaflets and cards promoting the survey. The
197 survey targeted people who live in or nearby the Kokemäenjoki River Basin or who visit
198 the area regularly. The main objective in the survey was to map people’s perceptions
199 about the important places related to the CES benefits in their daily landscape (Chan et
200 al., 2012; Haines-Young & Potschin, 2010). On the mapping page, respondents were able
201 to mark where they live and their important places as data points (Figure 2, accompanied
202 by the question, “What places are important for you in the Kokemäki region?”).
203 Respondents were able to map an unlimited number of important places and choose
204 related CES from a list of twenty items (e.g. moving in nature, hunting/fishing, doing
205 sports etc.) (Fig. 2). The CES list was guided by ecosystem service frameworks and
206 empirical studies applying participatory mapping (Haines-Young & Potschin, 2013; Mea,
207 2005; Vallés-Planells et al., 2014) and were adjusted to the local landscape context. Non-
208 spatial pages in the survey included questions about the demographic characteristics of
209 the respondents including e.g. gender, age and level of education. Data were collected
210 from 198 respondents. In addition to pre-defined options, respondents were also able to
211 describe in their own words why this place is important to them. Survey data were
212 collected and mapped homes treated anonymously in the analysis.



213

214 Figure 2. Screenshot from the survey showing the page for mapping important places and
 215 depicting the CES benefits linked to the mapped place.

216 **2.3 Land cover overlay**

217 The Densely Populated Areas dataset

218 (<http://metatieto.ymparisto.fi:8080/geoportal/catalog/search/resource/details.page?uui>

219 [d=%7B9085831B-8858-46E5-BAE1-3839CEC4CB52%7D](http://metatieto.ymparisto.fi:8080/geoportal/catalog/search/resource/details.page?uui)) and CLC 2018

220 (<http://metatieto.ymparisto.fi:8080/geoportal/catalog/search/resource/details.page?uui>

221 [d=%7B26EEEEBBB-FB5C-4045-B6DF-439F9B7D5C46%7D](http://metatieto.ymparisto.fi:8080/geoportal/catalog/search/resource/details.page?uui)) were downloaded from

222 the Finnish Environmental Institute open web service. The Density Populated Areas

223 dataset was used to distinguish mapped places between rural and urban areas. Based on

224 CLC data, the Kokemäenjoki area was divided into four infrastructure types, which differ

225 from each other in terms of their landscape characteristics and human activities. CLC was

226 reclassified into four components: grey infrastructure (urban fabric, industrial,

227 commercial and transport units, mine, dump and construction sites, CLC classes 111-133,

228 142), green infrastructure (forest areas, shrub and/or herbaceous vegetation associations

229 and urban green areas, CLC classes 141, 244–324 and 333), yellow infrastructure (arable

230 lands, permanent crops, pastures and heterogeneous agricultural areas, CLC classes 211–

231 243, 331, 332, 334 and 335) and blue infrastructure (inland wetlands, inland waters and
232 marine waters, CLC classes 411–523). We overlaid the respondents' important places to
233 this base map with further distinction between urban and rural areas to study the relation
234 between the landscape structure and perceived ecosystem services in both urban and rural
235 areas. All spatial analyses were done using ArcGIS Pro (Esri, 2018).

236 ***2.4 Spatial patterns of CES***

237 We created kernel density maps for both urban and rural ecosystem services to visualise
238 their spatial patterns and hot spots (Silverman, 1986). We applied 'expected mean
239 distance' to use it for kernel density estimation as a radius. The Average Nearest Neighbor
240 tool was used to measure the distance between each important point centroid and its
241 nearest neighbor's location. The tool averages all these nearest neighbor distances and
242 shows that they are clustered. The Nearest Neighbor Index is expressed as the ratio of the
243 Observed Mean Distance to the Expected Mean Distance. The expected distance is the
244 average distance between neighbours in a hypothetical random distribution. If the index
245 is less than 1, the pattern exhibits clustering; if greater than 1, the trend is towards
246 dispersion or competition (Scott & Janikas, 2009; ArcGIS, 2020). To use the same radius
247 for both urban and rural area density, we used the average value (urban value+rural
248 value/2) of the Expected Mean Distance for urban and rural, i.e. 2000-metre radius.

249 ***2.5 Statistical analysis of CES***

250 Respondents' profiles were analysed in SPSS 26 (IBM, 2019) through descriptive
251 analyses on age, gender and education. Perceived CES benefits were analysed by cross-
252 tabulation, chi square statistic and standardised residuals to examine the CES distribution
253 on each infrastructure's type in both urban and rural areas. We used cross tabulations to
254 compare the mapped CES benefits (categorical variables) with the urban and rural areas

255 and each infrastructure type. Standardised residuals were calculated to identify whether
 256 the number of CES points differed significantly from the expected number of points in
 257 both urban and rural and each infrastructure type. Expected counts are the projected point
 258 frequencies in each infrastructure type if the null hypothesis (h_0) is true (i.e. there is no
 259 association between CES in both urban and rural and infrastructure types). We further
 260 used the chi-square test to test whether mapped CES benefits differed significantly
 261 between urban and rural areas and between different infrastructure types, respectively. A
 262 value of 0.05 was used as a limit of statistical significance.

263 3. Results

264 3.1 Respondents' profile

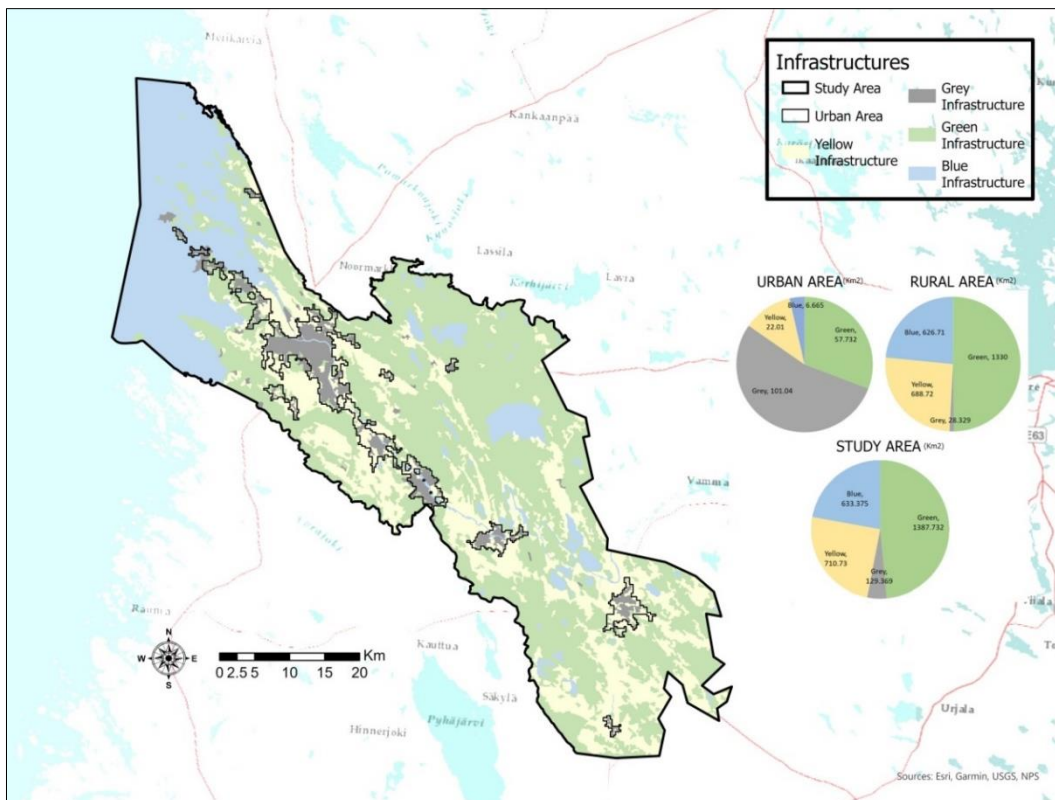
265 Women (50.3%) and men (48.7%) were nearly equally represented among the survey
 266 respondents (n=198). Further, 26.8% of respondents have a university or higher education
 267 degree, and 14.1% have post-secondary education degrees. There are 8% of respondents
 268 younger than 30 years, 64% between ages 30 and 59, and 28% are 60 years or older
 269 (n=188) (Table 1). Three respondents defined neither age nor gender. The answers to the
 270 question 'Where do you live?' were limited; 14 respondents indicated that they live in
 271 rural areas and 33 in urban areas.

272 Table 1 Subdivisions of respondents by age group and gender

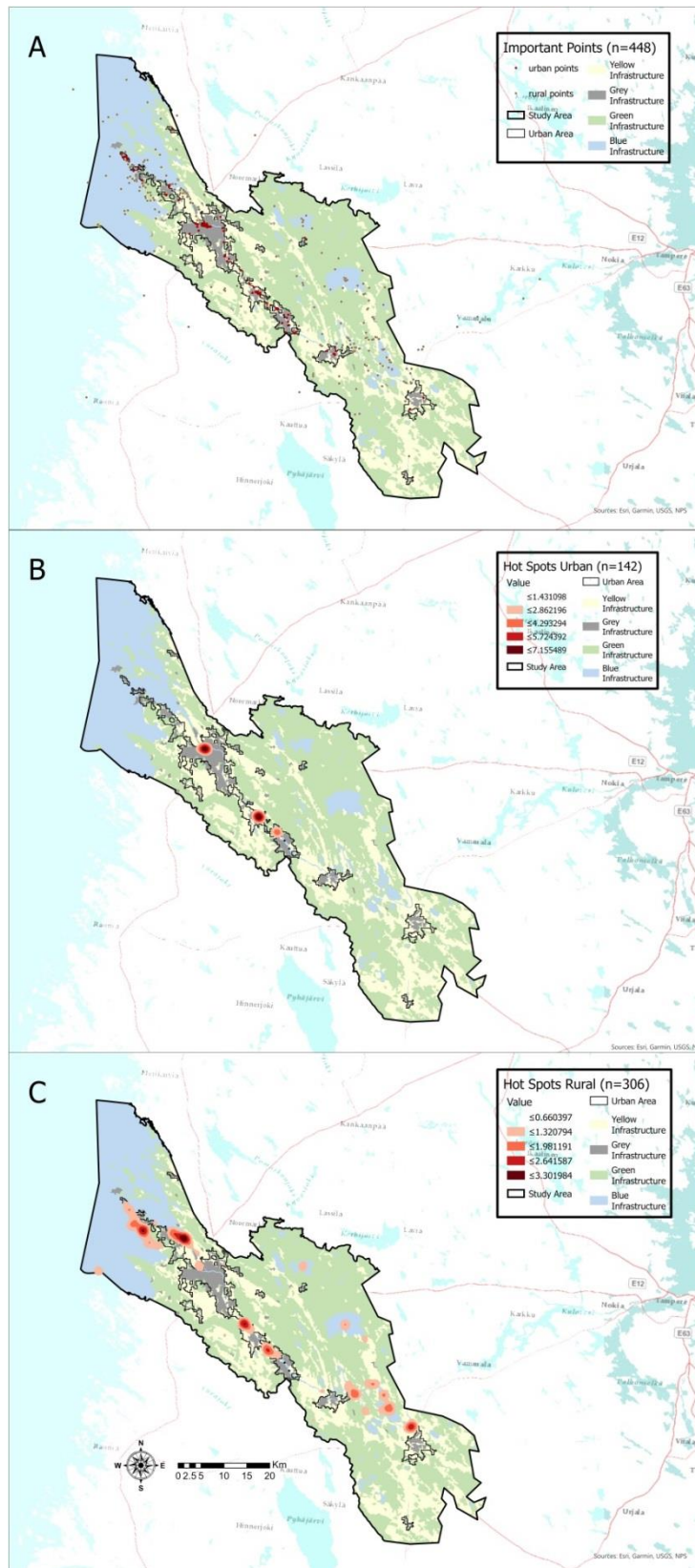
Age groups	Respondents	
	Female	Male
<18	1	1
18–25	3	5
26–35	7	15
36–45	20	16
46–55	28	27
56–65	26	18
66<	12	9
Non-defined	2	5
Total	99	96

273 **3.2 Infrastructure types in rural and urban areas and the spatial distribution of mapped**
274 **important places**

275 For the whole study area, the biggest infrastructure type is green, with 49% (1,387 km²).
276 For urban areas, the biggest infrastructure type is grey, with 45% (12,874.95 km²) of the
277 whole urban area. For rural area, the green infrastructure is biggest, with 49% (14,019.39
278 km²) (Figure 3).



279
280 Figure 3. Four types of infrastructure in urban and rural areas in the Kokemäenjoki area.
281 The respondents mapped 448 important places in total in the study area (Fig. 4A). Rural
282 areas (68.3%, 306 of mapped important places) (Fig. 4C) have the most prominent
283 hotspots of mapped places compared to urban areas (31.7%, 142 of mapped important
284 places) (Fig. 4B), where mapped CES benefits are mostly found in the grey infrastructure.



285

286 Figure 4. Spatial patterns of mapped important places in the Kokemäenjoki area. A)
 287 distribution of mapped important places, B) hot spots of mapped places in urban areas,
 288 and C) hot spots of mapped places in rural areas.

289 **3.3 Perceived CES benefits in different infrastructures**

290 In total, 1800 CES benefits were identified in connection to the 448 mapped important
291 places. In relation to each mapped place, on average, 2.9 (SD 1.1) different CES benefits
292 were identified. The mean number of perceived benefits was higher for green (4.22
293 benefits) and blue (4.07 benefits) infrastructures than for grey (3.89 benefits) and yellow
294 (3.81 benefits). CES benefits that were perceived in relation to the mapped places are
295 most commonly found in connection to the blue infrastructure (43%, 773 identified
296 benefits). In rural areas, we found 47% (602 identified benefits) of CES benefits on blue
297 infrastructure, 28% (359 identified benefits) on green infrastructure, 22% (279 identified
298 benefits) on yellow infrastructure and 3% (46 identified benefits) on grey infrastructure
299 (Figure 5). In urban areas, we found the majority of identified benefits on grey
300 infrastructure at 48% (246), blue infrastructure had 33% (171), yellow had 13% (68), and
301 green infrastructure had 6% (29) of identified benefits (Figure 6).

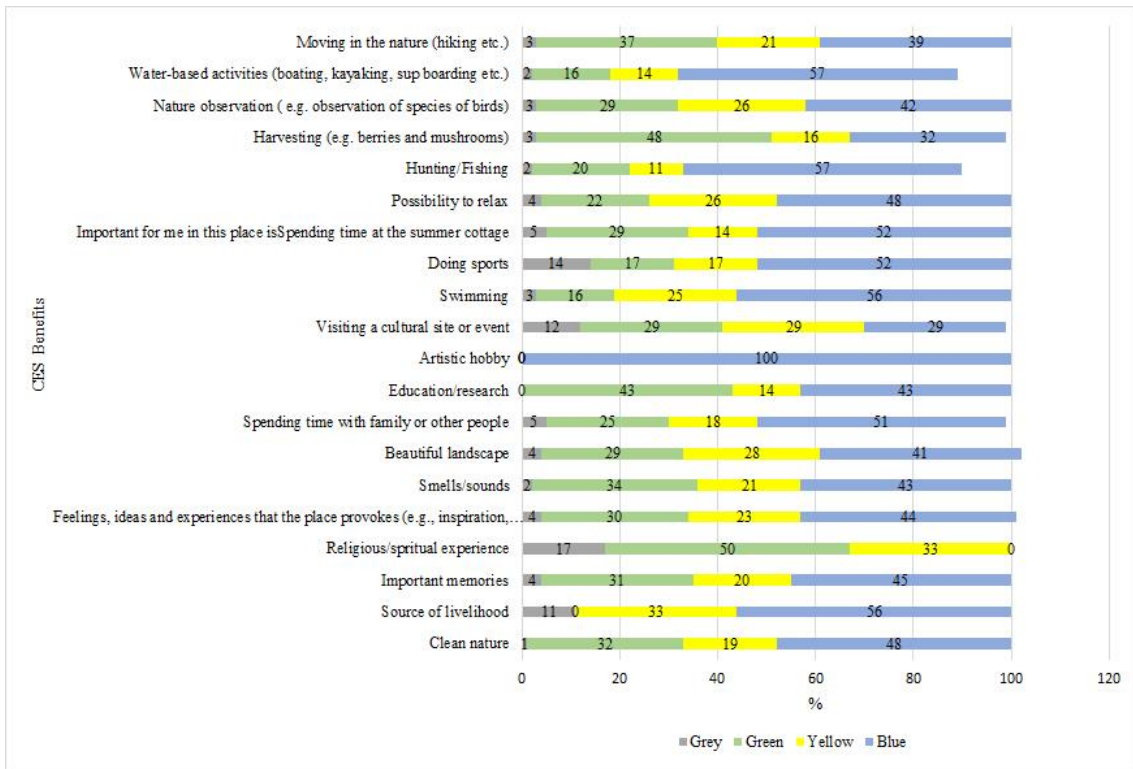


Figure 5. Perceived CES benefits on each infrastructure type in rural areas (%)

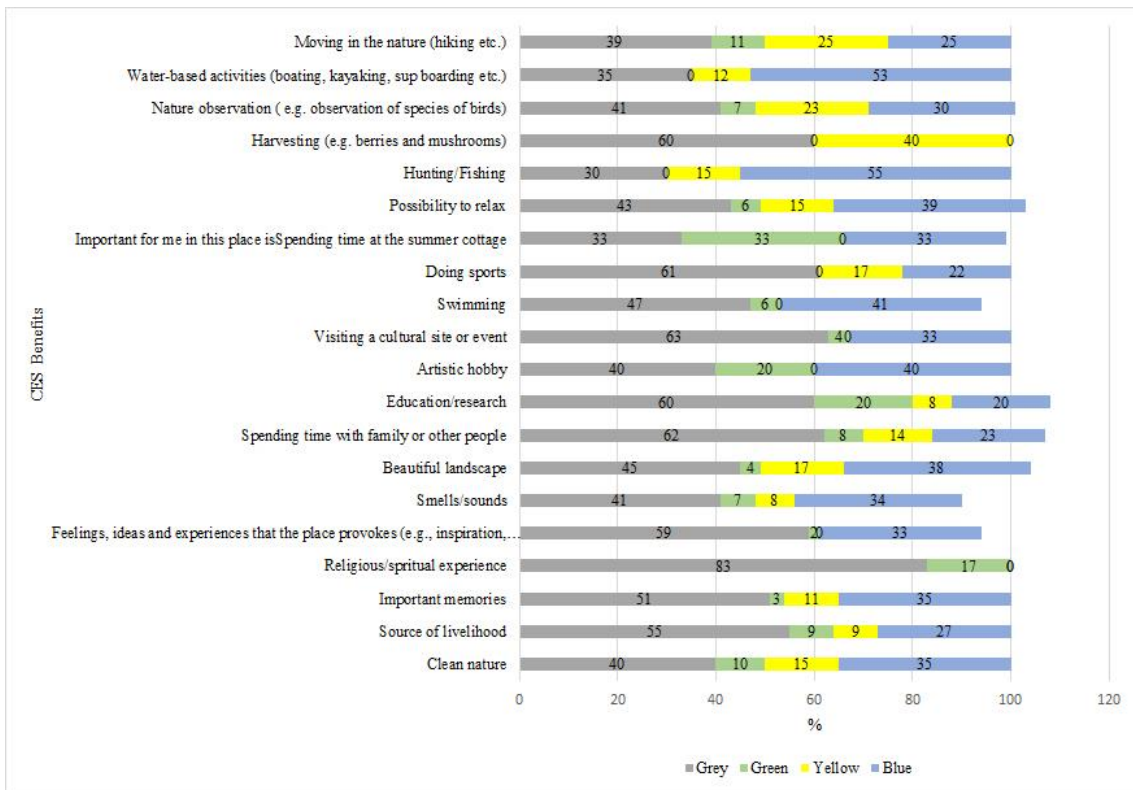


Figure 6. Perceived CES benefits on each infrastructure type in urban areas (%)

1 We found a statistically significant difference in CES perception among different
 2 infrastructures. Residents' preferences showed the dominance of blue infrastructure and
 3 statistically significant overrepresentation was observed for 8 of the 20 CES as being
 4 important to the participants (45.9%), namely moving in nature (35.5% of CES benefits
 5 in blue infrastructure), water-based activities (64.7%), nature observation (39.5%),
 6 hunting/fishing (64%), sport activities (40.4%), spending time with family or other people
 7 (41.7%), smells/sounds (41.4%), and clean nature (like clean air and fresh water) (Table
 8 2).

9 Statistically significant differences in CES perception were related to other
 10 infrastructures as follows: Harvesting is carried out by 41.7% in green infrastructure
 11 ($X^2(3, N = 36) = 11.4, p = .010$). Visiting a cultural site or an event is preferred by 43.2%
 12 in grey infrastructure ($X^2(3, N = 44) = 26, p = .000$). A place was considered important
 13 in terms of religious/spiritual experiences by 50% in grey infrastructure ($X^2(3, N = 12)$
 14 $= 14.6, p = .002$).

15 Table 2. Share (%) of CES benefits perceived in connection to important places in grey,
 16 green, yellow and blue infrastructures and the p-values from chi-square tests. Significant
 17 associations are bolded.

Cultural Ecosystem Services (CES) Benefits	Grey %	Green %	Yellow %	Blue %	p-value
Moving in nature (hiking etc.)	12.3	30.4	21.7	35.5	0.002
Water-based activities (boating, kayaking, sup boarding etc.)	7.8	13.7	13.7	64.7	0.000
Nature observation (e.g. observation of species of birds)	10.5	24.5	25.5	39.5	0.000
Harvesting (e.g. berries and mushrooms)	11.1	41.7	19.4	27.8	0.010
Hunting/Fishing	9.3	14.7	12.0	64.0	0.000
Possibility to relax	16.2	16.8	22.5	44.5	0.372
Important for me in this place is spending time at the summer cottage	8.3	29.2	12.5	50.0	0.310
Doing sports	31.9	10.6	17.0	40.4	0.013
Swimming	18.4	12.2	18.4	51.0	0.301
Visiting a cultural site or event	43.2	13.6	11.4	31.8	0.000

Artistic hobby	25.0	12.5	0.0	62.5	0.345
Education/research	25.0	33.3	8.3	33.3	0.426
Spending time with family or other people	24.3	19.1	14.8	41.7	0.017
Beautiful landscape	17.0	20.0	23.0	40.0	0.907
Smells/sounds	10.5	27.8	20.3	41.4	0.015
Feelings, ideas and experiences that the place provokes (e.g. inspiration, excitement)	22.7	19.7	17.4	40.2	0.080
Religious/spiritual experience	50.0	33.3	16.7	0.00	0.002
Important memories	20.4	21.3	16.7	41.7	0.335
Source of livelihood	35.0	5.0	20.0	40.0	0.078
Clean nature	7.4	28.7	18.0	45.9	0.000

18 4. Discussion

19 *4.1 CES benefits in urban and rural areas and land cover infrastructures*

20 This study analysed the CES benefits along the urban–rural gradient of the Kokemäenjoki
21 area in Southwest Finland. Participatory mapping and spatial analyses of CES have
22 broadened the understanding of the connections between people and their landscape for
23 both urban and rural areas. We have shown that perceptions of the landscape and their
24 links to CES by people familiar with the Kokemäenjoki area can be spatially and
25 statistically analysed. As only few studies have addressed ecosystem services along the
26 urban–rural gradient, our findings are relevant for providing a better understanding of
27 land cover and ecosystem services in terms of perceived CES benefits. Several studies
28 have applied PPGIS in the contexts of CES and different land covers (Brown &
29 Fagerholm, 2015; Brown & Hausner, 2017; Brown & Kytä, 2014b). Additionally, our
30 study analyses CES in relation to different infrastructures.

31 Our analyses are based on local and regional spatial data available for the
32 Kokemäenjoki area. This study does not aim to provide accurate data for the relations
33 between CES, infrastructures and different land cover types, e.g. urban and rural. Rather,
34 the main objective of this study was to show whether there is a significant association
35 between CES and infrastructures in rural and urban areas.

36 Our convenience sampling approach has the potential to cover all residents and
37 visitors of the area who can be reached through the contacts we made (i.e. press release
38 in municipalities, regional council actors, regional planners, nature conservation
39 association and social media). Hence, the potential crowd is much larger than we reached.
40 Nevertheless, the number of respondents allows our analysis to derive conclusions on the
41 CES provided by the study area, and we find that even with this limited number of
42 respondents, we can showcase the applicability of the ecosystem service concept in
43 landscape planning. However, we acknowledge that a random sampling is an essential
44 addition to this type of voluntary participation to reach more representative views (Brown
45 et al., 2014).

46 Using the Corine 2018 data, we demonstrated what kind of land covers are related
47 to the infrastructures and analysed the mapped perceptions of people concerning each of
48 the four types of infrastructure. We found some comparable results in other studies that
49 applied a similar methodology: Our findings proved that the analysed CES are spatially
50 clustered to hotspots (Garcia-Martin et al., 2017; Plieninger et al., 2013) and confirmed
51 that, in terms of providing CES, the importance of blue infrastructure (the sea and river
52 areas) is related to the recreational and aesthetic benefits (Andersson et al., 2015; Grizzetti
53 et al., 2016; Haase, 2015). The Kokemäenjoki area is a coastal site with some riverside
54 populated centres, and therefore, people perceive many CES benefits particularly in
55 relation to the water areas in both urban and rural areas. Attractive seashore and closeness
56 to the sea together with the settlement and several historical places located along the river
57 are major factors in why many important places are related to the blue infrastructure.
58 Green and yellow infrastructures are related to active and passive recreational benefits,
59 like doing sports and observing nature. This result is also similar to related studies
60 (Kabisch & Haase, 2014; Mell, 2010). On the other hand, it is interesting that grey

61 infrastructure is related to spiritual benefits, which could be due to the existence of
62 historical places along the river that are close to the major settlements. At this point, this
63 study showed that grey infrastructures are interesting in some respects but poor or
64 dangerous for nature if not combined with other types of infrastructures in terms of the
65 ecosystems' sustainability.

66 It is clearly seen that blue infrastructure is significantly different than the other
67 infrastructure types in terms of CES benefits. In rural areas, blue infrastructure provides
68 more CES than the other infrastructures, and the most provided CES in blue infrastructure
69 is 'artistic hobby' (hobbies related to art, such as painting, drawing, sketching, making
70 sculptures, writing poetry, dancing etc.). In urban areas, grey infrastructure provides more
71 CES than the other infrastructures, and the most provided CES benefits are
72 'religious/spiritual experiences'. Previous studies highlight the importance of green
73 infrastructure and assume that green infrastructures produce more ecosystem services in
74 both urban and rural areas (Di Marino et al., 2019; Lindley et al., 2018). However, we
75 found that blue infrastructure produced more ecosystem services in this landscape where
76 river and water elements are dominating in rural areas. Blue infrastructure was also
77 significant in urban areas, although grey infrastructure was related to most CES benefits
78 in urban areas related to the water elements, This may be unexpected because some
79 studies (Dong et al., 2017; Svendsen et al., 2012; Tiwary & Kumar, 2014) indicate that
80 green infrastructure in urban areas supports human well-being more than the other types
81 of infrastructure. It could be true for provisioning, regulating or supporting ecosystem
82 services; however, for CES like sense of place or spiritual values, grey infrastructure is
83 more preferred than the other types of infrastructures based on the findings of our study.

84 The results showed that activities related to recreation and cultural values are the
85 most perceived common values in both urban and rural areas and in different

86 infrastructures. Perceived values are significantly related to the presence of water, green
87 areas, and cultural heritage. However, respondents in urban and rural areas have their own
88 values connected to the infrastructures in terms of their social and cultural characteristics
89 (Garcia-Martin et al., 2017).

90 ***4.2 Implementation in landscape planning and management***

91 The ecosystem service concept became more predominant over the last decade in
92 landscape planning and management (Baró et al., 2017; De Groot et al., 2010; Frank et
93 al., 2012). The spatial assessment of ecosystem services could support the landscape
94 planning process towards sustainability. Additionally, analysing the infrastructures
95 spatially can be a way to understand the natural environment and relations between social
96 and ecological structures.

97 In Finland, ecosystem services are not embedded in the practical planning
98 processes at a regional level (Niemelä et al., 2010; Rinne & Primmer, 2016). Therefore,
99 the applicability of the ecosystem service concept was tested and discussed during the
100 SustainBaltic project in order to define different values of blue-green environments like
101 agricultural areas, forests and semi-natural areas, and water bodies. The recent literature
102 is quite familiar with the ecosystem services concept, but there is a lack of information
103 about the implementation of this concept into landscape planning and management. There
104 are, however, few examples from Finland related to its practical implementation from
105 recent years. Tammi et al. (2017) have discussed the challenges of ecosystem service
106 concept implementation in their paper. They indicated that there is indeed a need to find
107 a connection between knowledge of green and blue areas and ecosystem services hotspot
108 areas. In our study, we aimed to find a relationship between CES benefits and different
109 environments, which are called ‘infrastructures’ in this article. Our findings showed that
110 different infrastructures provide CES benefits at different levels.

111 The SustainBaltic project aimed to define the trends for sustainable marine and
112 coastal tourism and to introduce the ecosystem service concept in regional land use
113 planning. The project team found that increased communication and an exchange of local
114 views and values on the prevailing land–sea interactions was one important result of the
115 project.

116 Our findings confirmed that spatial assessment of the CES could be a key element
117 in analysing spatially clustered urban and rural areas and their impact on landscapes (Hou
118 et al., 2020). The ecosystem service concept in this context can be used as a platform to
119 promote more holistic consideration of values in different environments and people’s
120 preferences towards them. The vagueness of the concept, however, was considered one
121 of the biggest challenges in its practical application, and more explicit tools to describe
122 different ecosystem services were encouraged. The need for new tools is particularly
123 evident in CES assessments because these values are often linked to people’s personal
124 preferences rather than to explicit physical or ecological measures (Bagstad et al., 2017).

125 Cultural and natural landscape zones could be created from the CES hotspots in
126 both urban and rural areas. They could be a way of measuring CES as part of the land use
127 planning process and further of providing their sustainable management alongside other
128 ecosystem services. Van Riper et al. (2017) emphasised the growing interest in PPGIS,
129 particularly its ability to get knowledge from perceived values about biodiversity and its
130 potential to mix social and ecological data that may inform landscape management
131 decisions. Participatory mapping is also a helpful tool not only for measuring landscape
132 attractiveness and identifying the elements that compromise the landscape value (De
133 Vries et al., 2013) but also for building local communities’ cultural shared scenarios
134 (Assumma & Ventura, 2014).

135 In our study, we used PPGIS to link CES to explicit sites and infrastructures
136 within the study area. This kind of data can provide significant insights into the features
137 which are valued by the local people. These features need to be acknowledged in planning
138 to preserve social and cultural characteristics of the planning area and to avoid potential
139 conflicts between local stakeholders. If new activities are proposed for areas with high
140 numbers of CES benefits, conflicts with local stakeholders may arise, and this may delay
141 the entire planning process. Brown and Raymond (2014) evaluated three methods for
142 identifying potential land use conflict using mapped place data as a function of the level
143 of agreement on land use preferences and values of the place. If the PPGIS is collected
144 before the actual planning process, the sites with high CES benefits can be identified in
145 the early stages of the planning process and suitable actions can be taken either to avoid
146 sites that are prone to conflicts or to start discussions with the local stakeholders about
147 the impact mitigation.

148 **Funding:** This work was supported by the Interreg Central Baltic Programme 2014-2020
149 European Regional Development Fund (CB354 SustainBaltic project). E. Seda
150 ARSLAN's contribution was funded through the grant from The Scientific
151 and Research Council of Turkey (TUBİTAK), International Postdoctoral Research
152 Fellowship Program (2219).

153 **Acknowledgments:** We would like to thank the residents in the study area for
154 participating in the survey.

155 **Conflicts of Interest:** The authors declare no conflict of interest. The funders had no role
156 in the design of the study; in the collection, analyses, or interpretation of data; in the
157 writing of the manuscript, or in the decision to publish the results.

158 **References**

- 159 Ahtiainen, H., & Öhman, M. C. (2014). *Ecosystem services in the Baltic Sea*.
- 160 Andersson, E., Tengö, M., McPhearson, T., & Kremer, P. (2015). Cultural ecosystem
161 services as a gateway for improving urban sustainability. *Ecosystem Services*, *12*,
162 165–168.
- 163 ArcGIS, online. (n.d.). *ArcGIS Desktop*. Retrieved September 4, 2020, from
164 [https://desktop.arcgis.com/en/arcmap/10.3/tools/spatial-statistics-](https://desktop.arcgis.com/en/arcmap/10.3/tools/spatial-statistics-toolbox/average-nearest-neighbor.htm#ESRI_USAGES_AB280A56C7994327B1CB6A6E60DF4503)
165 [toolbox/average-nearest-](https://desktop.arcgis.com/en/arcmap/10.3/tools/spatial-statistics-toolbox/average-nearest-neighbor.htm#ESRI_USAGES_AB280A56C7994327B1CB6A6E60DF4503)
166 [neighbor.htm#ESRI_USAGES_AB280A56C7994327B1CB6A6E60DF4503](https://desktop.arcgis.com/en/arcmap/10.3/tools/spatial-statistics-toolbox/average-nearest-neighbor.htm#ESRI_USAGES_AB280A56C7994327B1CB6A6E60DF4503)
- 167 Bachi, L., Ribeiro, S. C., Hermes, J., & Saadi, A. (2020). Cultural Ecosystem Services
168 (CES) in landscapes with a tourist vocation: Mapping and modeling the physical
169 landscape components that bring benefits to people in a mountain tourist
170 destination in southeastern Brazil. *Tourism Management*, *77*, 104017.
- 171 Bagstad, K. J., Semmens, D. J., Ancona, Z. H., & Sherrouse, B. C. (2017). Evaluating
172 alternative methods for biophysical and cultural ecosystem services hotspot
173 mapping in natural resource planning. *Landscape Ecology*, *32*(1), 77–97.
- 174 Baró, F., Gómez-Baggethun, E., & Haase, D. (2017). Ecosystem service bundles along
175 the urban-rural gradient: Insights for landscape planning and management.
176 *Ecosystem Services*, *24*, 147–159.
- 177 Battisti, L., Corsini, F., Gusmerotti, N. M., & Larcher, F. (2019). Management and
178 perception of Metropolitan Natura 2000 Sites: A case study of La Mandria Park
179 (Turin, Italy). *Sustainability*, *11*(21), 6169. <https://doi.org/10.3390/su11216169>
- 180 Berghöfer, A., Mader, A., Patrickson, S., Calcaterra, E., Smit, J., Blignaut, J., de Wit, M.,
181 & van Zyl, H. (2011). TEEB Manual for cities: Ecosystem services in urban
182 management. *The Economics of Ecosystems and Biodiversity*, *Suiza*.

- 183 Bieling, C., & Plieninger, T. (2013). Recording manifestations of Cultural Ecosystem
184 Services in the landscape. *Landscape Research*, 38(5), 649–667.
185 <https://doi.org/10.1080/01426397.2012.691469>
- 186 Bonilla-Bedoya, S., Estrella, A., Yáñez, A. V., & Herrera, M. Á. (2020). Urban socio-
187 ecological dynamics: Applying the urban-rural gradient approach in a high
188 Andean city. *Landscape Research*, 45(3), 327–345.
189 <https://doi.org/10.1080/01426397.2019.1641589>
- 190 Brown, G. (2012). An empirical evaluation of the spatial accuracy of public participation
191 GIS (PPGIS) data. *Applied Geography*, 34, 289–294.
- 192 Brown, G. (2013). Relationships between spatial and non-spatial preferences and place-
193 based values in national forests. *Applied Geography*, 44, 1–11.
- 194 Brown, G., & Fagerholm, N. (2015). Empirical PPGIS/PGIS mapping of ecosystem
195 services: A review and evaluation. *Ecosystem Services*, 13, 119–133.
- 196 Brown, G., & Hausner, V. H. (2017). An empirical analysis of cultural ecosystem values
197 in coastal landscapes. *Ocean & Coastal Management*, 142, 49–60.
- 198 Brown, G., & Kyttä, M. (2014a). Key issues and research priorities for public
199 participation GIS (PPGIS): A synthesis based on empirical research. *Applied*
200 *Geography*, 46, 122–136.
- 201 Brown, G., & Kyttä, M. (2014b). Key issues and research priorities for public
202 participation GIS (PPGIS): A synthesis based on empirical research. *Applied*
203 *Geography*, 46, 122–136.
- 204 Brown, G., & Raymond, C. M. (2014). Methods for identifying land use conflict potential
205 using participatory mapping. *Landscape and Urban Planning*, 122, 196–208.
206 <https://doi.org/10.1016/j.landurbplan.2013.11.007>

- 207 Burkhard, B., Kroll, F., Müller, F., & Windhorst, W. (2009). Landscapes' capacities to
208 provide ecosystem services—a concept for land-cover based assessments.
209 *Landscape Online*, 15(1), 1–22.
- 210 Burkhard, B., Kroll, F., Nedkov, S., & Müller, F. (2012). Mapping ecosystem service
211 supply, demand and budgets. *Ecological Indicators*, 21, 17–29.
- 212 Burkhard, B., Santos-Martin, F., Nedkov, S., & Maes, J. (2018). An operational
213 framework for integrated Mapping and Assessment of Ecosystems and their
214 Services (MAES). *One Ecosystem*, 3 (2018).
- 215 Chan, K. M. A., Satterfield, T., & Goldstein, J. (2012). Rethinking ecosystem services to
216 better address and navigate cultural values. *Ecological Economics*, 74(C), 8–18.
217 <https://ideas.repec.org/a/eee/ecolec/v74y2012icp8-18.html>
- 218 Chen, X., de Vries, S., Assmuth, T., Dick, J., Hermans, T., Hertel, O., Jensen, A., Jones,
219 L., Kabisch, S., & Lanki, T. (2019). Research challenges for cultural ecosystem
220 services and public health in (peri-)urban environments. *Science of the Total
221 Environment*, 651, 2118–2129.
- 222 CoE, C. O. (2000). *European landscape convention*.
- 223 Costanza, R., d'Arge, R., De Groot, R., Farber, S., Grasso, M., Hannon, B., Limburg, K.,
224 Naeem, S., O'Neill, R. V., & Paruelo, J. (1997). The value of the world's
225 ecosystem services and natural capital. *Nature*, 387(6630), 253.
- 226 Daily, G. C. (1997). *Nature's services* (Vol. 19971). Island Press, Washington, DC.
- 227 De Groot, R. (2006). Function-analysis and valuation as a tool to assess land use conflicts
228 in planning for sustainable, multi-functional landscapes. *Landscape and Urban
229 Planning*, 75(3–4), 175–186.

- 230 De Groot, R. S., Alkemade, R., Braat, L., Hein, L., & Willemsen, L. (2010). Challenges
231 in integrating the concept of ecosystem services and values in landscape planning,
232 management and decision making. *Ecological Complexity*, 7(3), 260–272.
- 233 Di Marino, M., Tiitu, M., Lapintie, K., Viinikka, A., & Kopperoinen, L. (2019).
234 Integrating green infrastructure and ecosystem services in land use planning.
235 Results from two Finnish case studies. *Land Use Policy*, 82, 643–656.
236 <https://doi.org/10.1016/j.landusepol.2019.01.007>
- 237 Dipeolu, A. A., & Ibem, E. O. (2020). Green infrastructure quality and environmental
238 sustainability in residential neighbourhoods in Lagos, Nigeria. *International*
239 *Journal of Urban Sustainable Development*, 0(0), 1–16.
240 <https://doi.org/10.1080/19463138.2020.1719500>
- 241 Dong, X., Guo, H., & Zeng, S. (2017). Enhancing future resilience in urban drainage
242 system: Green versus grey infrastructure. *Water Research*, 124, 280–289.
243 <https://doi.org/10.1016/j.watres.2017.07.038>
- 244 Egoh, B. N., Reyers, B., Rouget, M., & Richardson, D. M. (2011). Identifying priority
245 areas for ecosystem service management in South African grasslands. *Journal of*
246 *Environmental Management*, 92(6), 1642–1650.
247 <https://doi.org/10.1016/j.jenvman.2011.01.019>
- 248 Eigenbrod, F. (2016). Redefining landscape structure for ecosystem services. *Current*
249 *Landscape Ecology Reports*, 1(2), 80–86.
- 250 Elwell, T. L., Gelcich, S., Gaines, S. D., & López-Carr, D. (2018). Using people's
251 perceptions of ecosystem services to guide modeling and management efforts.
252 *Science of the Total Environment*, 637, 1014–1025.

253 Ernoul, L., Wardell-Johnson, A., Willm, L., Béchet, A., Boutron, O., Mathevet, R.,
254 Arnassant, S., & Sandoz, A. (2018). Participatory mapping: Exploring landscape
255 values associated with an iconic species. *Applied Geography*, *95*, 71–78.

256 Fagerholm, N., Torralba, M., Moreno, G., Girardello, M., Herzog, F., Aviron, S., Burgess,
257 P., Crous-Duran, J., Ferreiro-Dominguez, N., Graves, A., Hartel, T., Macicasan,
258 V., Kay, S., Pantera, A., Varga, A., & Plieninger, T. (2019). Cross-site analysis
259 of perceived ecosystem service benefits in multifunctional landscapes. *Global*
260 *Environmental Change-Human and Policy Dimensions*, *56*, 134–147.
261 <https://doi.org/10.1016/j.gloenvcha.2019.04.002>

262 Fagerholm, N., Eilola, S., Kisanga, D., Arki, V., & Käyhkö, N. (2019). Place-based
263 landscape services and potential of participatory spatial planning in
264 multifunctional rural landscapes in Southern highlands, Tanzania. *Landscape*
265 *Ecology*, *34*(7), 1769–1787.

266 Fagerholm, N., Oteros-Rozas, E., Raymond, C. M., Torralba, M., Moreno, G., &
267 Plieninger, T. (2016). Assessing linkages between ecosystem services, land-use
268 and well-being in an agroforestry landscape using public participation GIS.
269 *Applied Geography*, *74*, 30–46.

270 Felipe-Lucia, M. R., & Comín, F. A. (2015). Ecosystem services–biodiversity
271 relationships depend on land use type in floodplain agroecosystems. *Land Use*
272 *Policy*, *46*, 201–210.

273 Frank, S., Fürst, C., Koschke, L., & Makeschin, F. (2012). A contribution towards a
274 transfer of the ecosystem service concept to landscape planning using landscape
275 metrics. *Ecological Indicators*, *21*, 30–38.

276 Garcia-Martin, M., Fagerholm, N., Bieling, C., Gounaridis, D., Kizos, T., Printsman, A.,
277 Müller, M., Lieskovský, J., & Plieninger, T. (2017). Participatory mapping of

278 landscape values in a Pan-European perspective. *Landscape Ecology*, 32(11),
279 2133–2150.

280 García-Nieto, A. P., Quintas-Soriano, C., García-Llorente, M., Palomo, I., Montes, C., &
281 Martín-López, B. (2015). Collaborative mapping of ecosystem services: The role
282 of stakeholders' profiles. *Ecosystem Services*, 13, 141–152.

283 Gebre, T., & Gebremedhin, B. (2019). The mutual benefits of promoting rural-urban
284 interdependence through linked ecosystem services. *Global Ecology and
285 Conservation*, 20. <https://doi.org/10.1016/j.gecco.2019.e00707>

286 Grizzetti, B., Lazanova, D., Liqueste, C., Reynaud, A., & Cardoso, A. C. (2016).
287 Assessing water ecosystem services for water resource management.
288 *Environmental Science & Policy*, 61, 194–203.

289 Haase, D. (2015). Reflections about blue ecosystem services in cities. *Sustainability of
290 Water Quality and Ecology*, 5, 77–83.

291 Haines-Young, R., & Potschin, M. (2010). The links between biodiversity, ecosystem
292 services and human well-being. *Ecosystem Ecology: A New Synthesis*, 1, 110–
293 139.

294 Haines-Young, R., & Potschin, M. (2013). Common International Classification of
295 Ecosystem Services (CICES): Consultation on Version 4, August–December
296 2012. EEA Framework Contract No EEA. *Contract No EEA/IEA/09/003*.

297 Hansen, R., & Pauleit, S. (2014). From multifunctionality to multiple ecosystem services?
298 A conceptual framework for multifunctionality in green infrastructure planning
299 for urban areas. *AMBIO*, 43(4), 516–529. [https://doi.org/10.1007/s13280-014-
300 0510-2](https://doi.org/10.1007/s13280-014-
300 0510-2)

301 Herlihy, P. H., & Knapp, G. (2003). Maps of, by, and for the peoples of Latin America.
302 *Human Organization*, 303–314.

303 Hou, L., Wu, F., & Xie, X. (2020). The spatial characteristics and relationships between
304 landscape pattern and ecosystem service value along an urban-rural gradient in
305 Xi'an City, China. *Ecological Indicators*, 108.
306 <https://doi.org/10.1016/j.ecolind.2019.105720>

307 IBM. (2019). *IBM SPSS Statistics for Windows*.

308 Jones, M. (2007). The European landscape convention and the question of public
309 participation. *Landscape Research*, 32(5), 613–633.
310 <https://doi.org/10.1080/01426390701552753>

311 Kabisch, N., & Haase, D. (2014). Green justice or just green? Provision of urban green
312 spaces in Berlin, Germany. *Landscape and Urban Planning*, 122, 129–139.

313 Kahila, M., Broberg, A., Kahila, F. M., & Broberg, A. (2017). Making cities wiser -
314 Crowdsourcing for better decisions. *FIG Monthly Article*.

315 Kahila-Tani, M., Kyttä, M., & Geertman, S. (2019). Does mapping improve public
316 participation? Exploring the pros and cons of using public participation GIS in
317 urban planning practices. *Landscape and Urban Planning*, 186, 45–55.

318 Kati, V., & Jari, N. (2016). Bottom-up thinking—Identifying socio-cultural values of
319 ecosystem services in local blue–green infrastructure planning in Helsinki,
320 Finland. *Land Use Policy*, 50, 537–547.

321 Kaymaz, I. (2013). Urban landscapes and identity. In *Advances in landscape architecture*.
322 IntechOpen.

323 Ko, H., & Son, Y. (2018). Perceptions of cultural ecosystem services in urban green
324 spaces: A case study in Gwacheon, Republic of Korea. *Ecological Indicators*, 91,
325 299–306.

326 Koschke, L., Fürst, C., Frank, S., & Makeschin, F. (2012). A multi-criteria approach for
327 an integrated land-cover-based assessment of ecosystem services provision to
328 support landscape planning. *Ecological Indicators*, *21*, 54–66.

329 Lamoureux, Z., & Fast, V. (2019). The tools of citizen science: An evaluation of map-
330 based crowdsourcing platforms. *Spatial Knowledge and Information Canada*,
331 *7*(4), 1.

332 Li, F., Liu, X., Zhang, X., Zhao, D., Liu, H., Zhou, C., & Wang, R. (2017). Urban
333 ecological infrastructure: An integrated network for ecosystem services and
334 sustainable urban systems. *Journal of Cleaner Production*, *163*, S12–S18.

335 Lilburne, L., Eger, A., Mudge, P., Ausseil, A.-G., Stevenson, B., Herzig, A., & Beare, M.
336 (2020). The Land Resource Circle: Supporting land-use decision making with an
337 ecosystem-service-based framework of soil functions. *Geoderma*, *363*, 114134.

338 Lin, B. B., Philpott, S. M., & Jha, S. (2015). The future of urban agriculture and
339 biodiversity-ecosystem services: Challenges and next steps. *Basic and Applied
340 Ecology*, *16*(3), 189–201.

341 Lindley, S., Pauleit, S., Yeshitela, K., Cilliers, S., & Shackleton, C. (2018). Rethinking
342 urban green infrastructure and ecosystem services from the perspective of sub-
343 Saharan African cities. *Landscape and Urban Planning*, *180*, 328–338.

344 Martín-López, B., Church, A., Başak Dessane, E., Berry, P., Chenu, C., Christie, M.,
345 Gerino, M., Keune, H., Osipova, E., Oteros-Rozas, E., Paillard, S., Rossberg, A.
346 G., Schröter, M., & Van Oudenhoven, A. P. E. (2018). *The IPBES regional
347 assessment report on biodiversity and ecosystem services for Europe and Central
348 Asia* (Chapter 2: Nature’s Contributions to People and Quality of Life).

349 Mea, M. E. A. (2005). Ecosystems and human well-being: Synthesis. *Island, Washington,*
350 *DC*.

351 Meacham, M., Queiroz, C., Norström, A. V., & Peterson, G. D. (2016). Social-ecological
352 drivers of multiple ecosystem services: What variables explain patterns of
353 ecosystem services across the Norrström drainage basin? *Ecology and Society*,
354 21(1).

355 Mell, I. C. (2010). *Green infrastructure: Concepts, perceptions and its use in spatial*
356 *planning*. Newcastle University.

357 Niemelä, J., Saarela, S.-R., Söderman, T., Kopperoinen, L., Yli-Pelkonen, V., Väre, S.,
358 & Kotze, D. J. (2010). Using the ecosystem services approach for better planning
359 and conservation of urban green spaces: A Finland case study. *Biodiversity and*
360 *Conservation*, 19(11), 3225–3243. <https://doi.org/10.1007/s10531-010-9888-8>

361 Palomo-Campesino, S., González, J. A., & García-Llorente, M. (2018). Exploring the
362 connections between agroecological practices and ecosystem services: A
363 systematic literature review. *Sustainability*, 10(12), 4339.

364 Plieninger, T., Dijks, S., Oteros-Rozas, E., & Bieling, C. (2013). Assessing, mapping, and
365 quantifying cultural ecosystem services at community level. *Land Use Policy*, 33,
366 118–129.

367 Pro, E. A. (2018). Windows. *ESRI: Redlands, CA, USA*.

368 Qiu, J., & Turner, M. G. (2013). Spatial interactions among ecosystem services in an
369 urbanizing agricultural watershed. *Proceedings of the National Academy of*
370 *Sciences*, 110(29), 12149–12154.

371 Rall, E., Bieling, C., Zytynska, S., & Haase, D. (2017). Exploring city-wide patterns of
372 cultural ecosystem service perceptions and use. *Ecological Indicators*, 77, 80–95.

373 Ramyar, R. (2019). Social–ecological mapping of urban landscapes: Challenges and
374 perspectives on ecosystem services in Mashhad, Iran. *Habitat International*, 92,
375 102043.

376 Reyers, B., Biggs, R., Cumming, G. S., Elmqvist, T., Hejnowicz, A. P., & Polasky, S.
377 (2013). Getting the measure of ecosystem services: A social–ecological approach.
378 *Frontiers in Ecology and the Environment*, 11(5), 268–273.

379 Rinne, J., & Primmer, E. (2016). A case study of ecosystem services in urban planning in
380 Finland: Benefits, rights and responsibilities. *Journal of Environmental Policy &*
381 *Planning*, 18(3), 286–305. <https://doi.org/10.1080/1523908X.2015.1076721>

382 Sieber, R. (2006). Public participation geographic information systems: A literature
383 review and framework. *Annals of the Association of American Geographers*,
384 96(3), 491–507.

385 Silverman, B. W. (1986). *Density estimation for statistics and data analysis* (Vol. 26).
386 CRC Press.

387 Small, N., Munday, M., & Durance, I. (2017). The challenge of valuing ecosystem
388 services that have no material benefits. *Global Environmental Change*, 44, 57–
389 67.

390 Soto-Montes-de-Oca, G., Bark, R., & González-Arellano, S. (2020). Incorporating the
391 insurance value of peri-urban ecosystem services into natural hazard policies and
392 insurance products: Insights from Mexico. *Ecological Economics*, 169, 106510.

393 Svendsen, E., Northridge, M. E., & Metcalf, S. S. (2012). Integrating grey and green
394 infrastructure to improve the health and well-being of urban populations. *Cities*
395 *and the Environment (CATE)*, 5(1), 3.

396 Tammi, I., Mustajärvi, K., & Rasinmäki, J. (2017). Integrating spatial valuation of
397 ecosystem services into regional planning and development. *Ecosystem Services*,
398 26, 329–344.

399 Tiwary, A., & Kumar, P. (2014). Impact evaluation of green–grey infrastructure
400 interaction on built-space integrity: An emerging perspective to urban ecosystem
401 service. *Science of the Total Environment*, 487, 350–360.

402 Vallés-Planells, M., Galiana, F., & Van Eetvelde, V. (2014). A classification of landscape
403 services to support local landscape planning. *Ecology and Society*, 19(1).

404 Villamor, G. B., Palomo, I., Santiago, C. A. L., Oteros-Rozas, E., & Hill, J. (2014).
405 Assessing stakeholders’ perceptions and values towards social-ecological systems
406 using participatory methods. *Ecological Processes*, 3(1), 22.

407 Von Braun, J. (2007). *Rural-urban linkages for growth, employment, and poverty*
408 *reduction* (pp. 7–9).

409 Zhang, S., & Ramírez, F. M. (2019). Assessing and mapping ecosystem services to
410 support urban green infrastructure: The case of Barcelona, Spain. *Cities*, 92, 59–
411 70.

412 Zhu, Y.-G., Reid, B. J., Meharg, A. A., Banwart, S. A., & Fu, B.-J. (2017). Optimizing
413 Peri-URban Ecosystems (PURE) to re-couple urban-rural symbiosis. *Science of*
414 *the Total Environment*, 586, 1085–1090.

415